Deciphering probabilistic species interaction networks

Francis Banville ^{1,2,3} Tanya Strydom ^{1,3} Penelope S. A. Blyth ⁴ Chris Brimacombe ⁵ Michael Catchen ^{3,6}
Gabriel Dansereau ^{1,3} Gracielle Higino ² Thomas Malpas ⁴ Hana Mayall ⁴ Kari Norman ¹
Dominique Gravel ^{2,3} Timothée Poisot ^{1,3}

¹ Université de Montréal ² Université de Sherbrooke ³ Quebec Centre for Biodiversity Science ⁴ University of Sheffield ⁵ University of Toronto ⁶ McGill University

Correspondance to:

Francis Banville — francis.banville@umontreal.ca

Representing species interactions probabilistically (how likely are they to occur?) as opposed to deterministically (are they occurring?) conveys uncertainties in our knowledge of interactions and information on their variability. The sources of uncertainty captured by interaction probabilities depend on the method used to evaluate them: uncertainty of predictive models, subjective assessment of experts, or empirical measurement of interaction spatiotemporal variability. However, guidelines for the estimation and documentation of probabilistic interaction data are lacking. This is concerning because our understanding and analysis of interaction probabilities depend on their sometimes elusive definition and uncertainty sources. We review how probabilistic interactions are defined at different spatial scales, from local interactions to regional networks (metawebs), with a strong emphasis on host-parasite and trophic (predatory and herbivory) interactions. These definitions are based on the distinction between the realization of an interaction at a specific time and space (local) and its biological or ecological feasibility (regional). Using host-parasite interactions in Europe, we illustrate how these two network representations differ in their statistical properties, specifically: how local networks and metawebs differ in their spatial and temporal scaling of probabilistic interactions, but not in their taxonomic scaling. We present two approaches to inferring binary interactions from probabilistic ones that account for these differences and show that systematic biases arise when directly inferring local networks from metawebs. Our results underscore the importance of more rigorous descriptions of probabilistic species interaction networks that specify their type of interaction (local or regional), conditional variables and uncertainty sources.

Last revision: December 16, 2024

1 Introduction

² Species interactions are variable and uncertain

As we try to navigate global biodiversity change, filling in knowledge gaps about biodiversity becomes 3 instrumental to monitoring and mitigating those changes (Abrego et al. 2021; Gonzalez & Londoño 2022; Hortal et al. 2015). However, cataloging species, populations and, in particular, ecological interactions (e.g., 5 predation, parasitism, and pollination) is a substantial challenge (Pascual et al. 2006; Polis 1991). There are 6 methodological and biological constraints that hinder our ability to document species interactions, inevitably 7 leading to uncertainty in our knowledge of interactions. For example, the spatial and temporal uncoupling of species (e.g., nocturnal and diurnal species coexisting in the same space with different daily activity timings, 9 Jordano 1987) and the large number of rare and cryptic interactions in a community, contribute to these 10 knowledge gaps by making it more difficult to observe interactions (Jordano 2016). 11 Several conditions must be satisfied for an interaction to be observed locally. First, both species must have 12 overlapping geographic ranges, i.e. they must co-occur within the region of interest (Cazelles et al. 2016; 13 Morales-Castilla et al. 2015). Second, they must have some probability of meeting within a defined time frame 14 (Poisot *et al.* 2015). Probabilities of interspecific encounters are typically low, especially for rare species with 15 low abundances (Canard et al. 2012; Canard et al. 2014; Vázquez et al. 2007). The probability that species 16 meet also depends on their biology, such as their phenology (Olesen et al. 2010; Singer & McBride 2012) and 17 discoverability (Broom & Ruxton 2005). Finally, when species do come into contact, an interaction occurs only 18 if their traits, such as their phenotypes (Bolnick et al. 2011; Gravel et al. 2013; Stouffer et al. 2011) and 19 behavior (Choh et al. 2012; Pulliam 1974), are locally compatible in that specific environment (Poisot et al. 20

21 2015). Because these conditions are not consistently met locally, there will inevitably be instances where

²² interactions will be observed and others where they will not.

Documenting the location and timing of interactions is difficult when accounting for the spatiotemporal
variability of ecological interactions (Poisot *et al.* 2012, 2015). Knowing the biological capacity of two species
to interact directly (via e.g., trophic interactions) is necessary but not sufficient for inferring their interaction at a
specific time and space. Environmental factors, such as temperature (Angilletta *et al.* 2004), drought
(Woodward *et al.* 2012), climate change (Araujo *et al.* 2011; Gilman *et al.* 2010; Woodward *et al.* 2010),
habitat characteristics (e.g., presence of refuges where prey can hide from predators, Grabowski 2004), and land
use change (Tylianakis *et al.* 2007), contribute to this spatiotemporal variability by impacting species

³⁰ abundance and traits. Interactions may also be influenced by a third species (e.g., a more profitable prey species,
³¹ Golubski & Abrams 2011; Sanders & van Veen 2012). Even under favorable circumstances, there remains a
³² possibility that the interaction does not occur locally, either due to the intricate nature of the system or simply
³³ by chance. If it does occur, it might go undetected, particularly if it happens infrequently. In this context, it is
³⁴ unsurprising that our knowledge of ecological interactions remains limited (Hortal *et al.* 2015) despite
³⁵ extensive biodiversity data collection (Schmeller *et al.* 2015).

We distinguish the variability of interactions from their uncertainty. Interaction variability is defined as the 36 changes in the occurrence or strength of interactions along spatial, temporal, or environmental axes (Poisot et al. 37 2015). It is a property of interactions that should be quantified if we aim for a comprehensive understanding of 38 ecological networks. Stochasticity is the inherent randomness or unpredictability of interactions that lead to this 39 variability. Conversely, uncertainty is defined as a lack of knowledge about the occurrence of interactions. 40 When using statistical models to infer interactions, uncertainty sources include input data, parameter, and model 41 structure uncertainties (Simmonds et al. 2024). Input data uncertainty arises from our inability to empirically 42 observe all interactions and from measurement errors in environmental and biological variables used for 43 inference. Parameter uncertainty represents a plausible range of values for a parameter whose exact value is 44 unknown. For example, we may calculate a range of plausible values for interaction variability (e.g., there could 45 be a 50% certainty that an interaction occurs 50% of the time). Model structure uncertainty recognizes that 46 different statistical models may adequately predict interactions. In contrast to variability, uncertainty can be 47 reduced by sampling additional data (except for model uncertainty, which will persist regardless of sampling 48 effort). Simmonds et al. (2024) underscores the importance of quantifying and reporting these diverse sources 49 of uncertainty, alongside ensuring their appropriate propagation to model output (such as predicted interactions) 50 and higher-level measures (such as network structure). While recognizing that these definitions may not be 51 universally accepted, clarifying the distinction between variability and uncertainty enables us to better 52 comprehend the sources of our knowledge gaps about ecological interactions. 53

54 Species interactions as probabilistic objects

The recognition of the intrinsic variability and uncertainty of species interactions has led ecologists to expand their representation of ecological networks to include a probabilistic view of interactions (Dallas *et al.* 2017; Fu *et al.* 2021; Poisot *et al.* 2016). This allows filling in the Eltonian shortfall (i.e., the gap between our current knowledge and a comprehensive understanding of interactions, Hortal *et al.* 2015) by modeling the probability

of occurrence of interactions (e.g., Gravel et al. 2019), which can be an important tool for directing efforts and 59 taking action (Carlson *et al.* 2021), especially in places where access and resources for research are scarce. A 60 probability is a measure of how likely a specific outcome is, based on both the uncertainty and variability of 61 interactions. Interaction probabilities may be uncertain when there is a distribution of plausible probability 62 values. The probabilistic representation of interactions has been applied to direct interactions, which are 63 conceptually and mathematically analogous regardless of their biological type (e.g., predation and pollination). 64 This is in contrast with indirect interactions (e.g., interspecific competition), which arise from distinct 65 ecological processes and are often not directly observable (Kéfi et al. 2015, 2016). By accounting for the 66 uncertainty and variability of direct interactions, networks of probabilistic interactions (which differ from 67 probabilistic networks describing the uncertainty and variability of the whole network) may provide a more 68 realistic portrait of species interactions. 69

Probabilistic interactions differ from binary interactions. Networks of probabilistic interactions, within a 70 Bayesian perspective, express our degree of belief (or confidence) regarding the occurrence or observation of 71 interactions. In a frequentist approach, they represent the expected relative frequencies of interactions over 72 many repeated trials or sampling events. In contrast, interactions are simply regarded as either occurring or not 73 in networks of deterministic binary interactions. Based on the scale at which they are estimated, interaction 74 probabilities may reflect our level of confidence in whether interactions will be observed, realized locally, or 75 biologically feasible. Our level of confidence should be more definitive (approaching either 0 or 1) as we extend 76 our sampling to a broader area and over a longer duration, thereby diminishing the uncertainty of our knowledge 77 of interactions (but not necessarily the estimation of their variability). In the broadest sense, binary interactions 78 are also a type of probabilistic interaction, in which the numerical value of an interaction is restrained to 0 79 (non-occurring) or 1 (occurring). In networks of probabilistic interactions, only forbidden interactions (i.e., 80 interactions prohibited by biological traits or species absence, Jordano et al. 2003; Olesen et al. 2010) have a 81 probability value of zero, provided that intraspecific trait variability is considered (Gonzalez-Varo & Traveset 82 2016). Understanding the nuances between probabilistic and binary interactions is essential for accurately 83 modeling and interpreting ecological networks. 84

The application and development of computational methods in network ecology, often based on a probabilistic representation of interactions, can alleviate (and guide) the sampling efforts required to document species interactions (Strydom *et al.* 2021). For example, statistical models can be used to estimate the uncertainty of pairwise interactions (Cirtwill *et al.* 2019) and the probability of missing (false negatives) and spurious (false

positives) interactions (Guimerà & Sales-Pardo 2009), helping us identify places where sampling is most 89 needed to reduce this uncertainty. Statistical models can also predict networks without prior knowledge of 90 pairwise interactions. They may do so using body size (Caron et al. 2024; Gravel et al. 2013; Petchey et al. 91 2008), phylogeny (Elmasri et al. 2020; Strydom et al. 2022), or a combination of niche and neutral processes 92 (Bartomeus et al. 2016; Pomeranz et al. 2019) for inference. Before being used to test ecological hypotheses, 93 predicted networks must be validated against empirical data (Brimacombe et al. 2024), which could be sampled 94 strategically to optimize the validation process. Topological null models, which generate networks of 95 probabilistic interactions by preserving chosen characteristics of the adjacency matrix of binary interactions 96 while intentionally omitting others (Bascompte et al. 2003; Fortuna & Bascompte 2006), are examples of 97 common probabilistic interaction models. Null models can produce underlying distributions of network 98 measures for null hypothesis significance testing. However, how the uncertainty of pairwise interactions 99 propagates to network structure (i.e., community-level properties driving the functioning, dynamics, and 100 resilience of ecosystems, McCann 2007; McCann 2011; Proulx et al. 2005; Rooney & McCann 2012) remains 101 to be elucidated. Many measures have been developed to describe the structure (Poisot et al. 2016) and 102 diversity (Godsoe et al. 2022; Ohlmann et al. 2019) of probabilistic interaction networks. These models and 103 measures support the use of probabilistic interactions for the study of a wide range of ecological questions, from 104 making better predictions of species distributions (Cazelles et al. 2016) to forecasting the impact of climate 105 change on ecological networks (Gilman et al. 2010). 106

¹⁰⁷ We lack a clear understanding of probabilistic species interactions

We still lack a precise definition of probabilistic interactions, which makes the estimation and use of these data more difficult. In this manuscript, we aim to take a step back by outlining different ways in which probabilistic interactions are defined and used in network ecology. We distinguish two broad categories of probabilistic interaction networks that necessitate distinct approaches: local networks describing probabilities of realized interactions, and regional networks (metawebs) describing probabilities of potential interactions. We highlight the distinctions in the ecological meaning of these two representations of interactions and examine their properties and relationships (particularly with space, time, and between each other).

The lack of clear guidelines on the use of probabilistic interaction data is worrisome, as it affects both data producers and re-users who generate and manipulate these numbers. This is concerning because sampling strategies and decisions regarding network construction can affect our understanding of network properties

(Brimacombe et al. 2023). There is currently no reporting standard that could guide the documentation of all 118 types of probabilistic interactions (Salim et al. 2022 discuss data standards for deterministic mutualistic 119 networks). Clear reporting standards for probabilistic interactions would support more adequate manipulation 120 and integration of interaction data from different sources and guard against possible misinterpretations arising 121 from ambiguous definitions of probabilistic interaction networks. Data documentation should outline the nature 122 (i.e., local or regional) and type (e.g., predatory or pollination) of interactions, provide information regarding 123 the taxonomic level, identities, and characteristics (e.g., life stages) of the individuals involved in an interaction, 124 present the mathematical formulation of probabilities, including clearly identified conditional variables (e.g., 125 spatial and temporal scales), and describe the methods and contexts (e.g., location, time, environmental 126 conditions) in which interactions were estimated. Inadequately documented probabilistic interaction data should 127 be used with caution when analyzing ecological networks. These broad principles remain relevant and 128 applicable across different types of direct interactions. In the following sections, we discuss the definitions, 129 conditions, and estimation of probabilistic interactions as we scale up from pairwise interactions to interactions 130 within local and regional networks. 131

¹³² Pairwise interactions: the building blocks of ecological networks

133 What are probabilistic interactions?

Consider a scenario where an avian predator has just established itself in a northern habitat home to a small 134 rodent. Suppose their interaction has not been previously observed, either because these species have never 135 co-occurred before or because previous sampling failed to detect an interaction despite their co-occurrence. 136 What is the probability that the rodent is part of the diet of the avian predator, or put differently, what is the 137 probability that they interact? Answering this question requires some clarification, as there are multiple ways to 138 interpret and calculate interaction probabilities. We could calculate the probability that the traits of these 139 species match, i.e. that the avian predator possesses the biological attributes to capture and consume the rodent. 140 We could also calculate the probability that their traits support an interaction under the typical environmental 141 conditions of the new habitat. For example, because avian predators hunt by sight, predation could be possible 142 in the absence of snow but highly improbable when snow is present, as rodents may use it as a shelter to hide 143 from predators. Finally, we could calculate the probability that the avian predator will consume the rodent at 144 *that* particular location, for which the spatial and temporal boundaries need to be specified. The estimation of 145

the probability of interaction between these two species, whether through predictive models or informative prior
 probabilities, hinges on our understanding of these probabilities and the specific ecological processes we aim to
 capture.

An important aspect to consider when estimating or using interaction probabilities is knowing if they describe 149 the probability of potential or realized interactions, as these two types of interactions have distinct meanings and 150 sources of uncertainty and variability. A potential (regional) interaction is defined as the biological or 151 ecological capacity of two taxa to interact (i.e., the probability that they interact if they were to encounter each 152 other, given sufficient time and appropriate environmental conditions) whereas a realized (local) interaction is 153 the occurrence or observation of this interaction in a well-defined space and time (i.e., the probability that they 154 interact locally). For two co-occurring taxa and over enough time, the probability of local interaction tends 155 toward the probability of regional (potential) interaction. A longer duration increases the probability that 156 species will eventually encounter each other and that local environmental conditions supporting an interaction 157 will occur, provided that species have the biological capacity to interact. Recognizing the distinction between 158 probabilistic regional and local interactions is crucial for accurately interpreting interaction probabilities in 159 ecological networks. 160

We use the terms *metaweb* (Dunne 2006) to designate regional networks of potential interactions and *local* 161 networks (Poisot et al. 2012) for those of realized interactions. Metawebs are the network analogs of the species 162 pool, where local networks originate from a subset of both species (nodes) and interactions (edges) of the 163 regional metaweb (Saravia et al. 2022). Without clear documentation, it can be challenging to know if published 164 probabilistic interaction networks describe local or regional interactions. When probabilistic local interactions 165 are used and interpreted incorrectly as regional interactions (and conversely), this may generate misleading 166 findings during data analysis. A better understanding of probabilistic local and regional interaction networks 167 would facilitate a more adequate use of interaction data (e.g., when studying network-area relationships in local 168 networks and metawebs) and prevent misinterpretations of the biological meaning of probabilistic interactions. 169

170 What is the outcome of probabilistic interactions?

171 The outcome of probabilistic interactions is usually binary

Local networks and metawebs, like any type of network, are made of nodes and edges that may be represented at different levels of organization. The basic units of ecological networks are individuals that interact with each other (e.g., by predation in food webs, Elton 2001), forming individual-based networks (Melián *et al.* 2011).
The aggregation of these individuals into more or less homogeneous groups (e.g., populations, species, families,
feeding guilds) allows us to represent nodes at broader taxonomic scales, which affects our interpretation of the
properties of these systems (Guimarães 2020; Hemprich-Bennett *et al.* 2021).

Ecologists have traditionally represented interactions (edges) as binary objects that were considered realized 178 after observing at least one individual from group *i* interact with at least another individual from group *j*. In an 179 adjacency matrix B of binary interactions, the presence or absence of an interaction $B_{i,j}$ between two taxa can 180 be viewed as the result of a Bernoulli trial $B_{i,j} \sim \text{Bernoulli}(\phi)$, with $\phi = P(B_{i,j} = 1)$ being the probability of 181 interaction. This interaction probability characterizes our limited ecological knowledge and/or the intrinsic 182 spatiotemporal variability of the interaction. It may be estimated through predictive models (e.g., those based 183 on biological traits and species abundances) or expert (prior) knowledge about the interaction. In networks of 184 probabilistic interactions, the edge values $P(B_{i,j} = 1)$ (which we denote as $P(B_{i,j})$ for simplicity and better 185 readability) are probabilistic events whose only two possible outcomes are the presence $(B_{i,j} = 1)$ or absence 186 $(B_{i,j} = 0)$ of an interaction between each pair of nodes. Depending on the type of probabilistic interaction 187 network (local network or metaweb), the mathematical formulation and interpretation of stochastic parameters 188 like $P(B_{i,j})$ can be linked to environmental and biological factors such as species abundances, species traits, 189 area, and time, for example using logistic regression with continuous explanatory variables. This allows us to 190 model the probability that at least two individuals interact under these conditions. 191

The variability of an interaction determines the number of networks in which it occurs. This number can be 192 predicted by using a Binomial distribution, assuming a constant interaction probability and independence 193 between interactions in different networks (trials). When considering uncertainties around the estimation of 194 $P(B_{i,i})$, a Beta distribution may be used to represent the relative likelihood of different probability values. For 195 example, when calculating the probability of interaction between two taxa based on their local abundances, any 196 uncertainty in their abundances would introduce uncertainty in the interaction probability at the local scale. If 197 we take into account the uncertainty of the interaction probability, a Beta-Binomial distribution can be used to 198 predict the number of networks in which the interaction occurs. Empirically observing an interaction between 199 two taxa at a given location and time provides important information that can be used to update previous 200 estimates of $P(B_{i,j})$, informing us of the conditions that enabled them to interact locally. By sampling binary 201 interactions in different contexts, we can thus estimate their local variability more precisely. 202

²⁰³ The outcome of probabilistic interactions may also be quantitative

Even though binary interaction networks constitute a highly valuable source of ecological information (Pascual 204 et al. 2006), they overlook interaction strengths. Represented in a quantitative adjacency matrix W, interaction 205 strengths better describe the energy flows, demographic impacts or frequencies of interactions between nodes 206 (Berlow *et al.* 2004; Borrett & Scharler 2019), with $W_{i,j}$ being a natural \mathbb{N} or real \mathbb{R} number depending on the 207 measure. For example, they may represent local interaction rates (e.g., the flower-visiting rates of pollinators in 208 a mutualistic network, Herrera 1989). Relative frequencies of interactions may be used as a measure of both the 209 strength and probability of local interactions that are biologically feasible. When interaction strengths 210 characterize predation pressure on prey, they can serve as parameters in a Lotka-Volterra model (e.g., 211 Emmerson & Raffaelli 2004). The extra amount of ecological information in quantitative networks typically 212 comes at a cost of greater sampling effort and data volume (Strydom et al. 2021), especially when using 213 predictive models that quantify the uncertainty and variability of quantitative interactions (Berlow et al. 2004). 214 However, if two taxa are repeatedly found together without interacting, there may be more uncertainty about 215 their capacity to interact than their interaction strength (which would assuredly be close to 0). 216

Like binary interaction networks, the uncertainty and variability of interaction strengths can be represented 217 probabilistically. Interaction strengths can follow many probability distributions depending on the measure. For 218 instance, they can follow a Poisson distribution $W_{i,i} \sim \text{Poisson}(\lambda_{i,i}t_0)$ when predicting the number of 219 interactions between individuals during a time interval t_0 , with $\lambda_{i,j}$ being the expected rate at which individuals 220 of taxa *i* and *j* interact (e.g., the expected number of prey *j* consumed by all predators *i*). The Poisson 221 distribution can also be 0-inflated when taking into account non-interacting taxa (e.g., Boulangeat et al. 2012 222 employ a 0-inflated model to analyze species abundance following the modeling of species presence and 223 absence), which constitute the majority of taxa pairs in most local networks (Jordano 2016). Regardless of the 224 measure, estimating the uncertainty of quantitative interactions enables us to consider a range of possible values 225 of interaction strength. 226

Because of the methodological difficulties typically encountered when building deterministic quantitative
networks, binary interaction networks, which are usually easier to sample (Jordano 2016) and predict (Strydom *et al.* 2021), have been more frequently studied and used. Mathematical models such as Ecopath (Plagányi &
Butterworth 2004) partially mitigate these difficulties, but the number of biological parameters required to make
predictions hinders their application in many systems. Moreover, most published probabilistic interaction

networks (e.g., Strydom *et al.* 2022) and methods (e.g., Poisot *et al.* 2016) involve probabilistic interactions
whose outcome is binary. This underlines the need for better guidelines on the interpretation and manipulation
of probabilistic interactions with binary outcomes first, to ensure the appropriate use of these networks and
methods. For these reasons, the primary focus of the remainder of this manuscript is on the interpretation of
interaction probabilities that determine the presence or absence of interactions, in both local networks and
metawebs.

²³⁸ Local networks: communities interacting in space and time

239 What are local probabilistic interactions?

Local networks of probabilistic interactions describe how likely taxa are to interact in a local context. Local 240 interactions are contingent upon the environmental conditions experienced by the community and the matching 24 of taxa's local biological traits. In local networks, edges commonly represent our degree of belief that two taxa 242 interact in nature, but can also represent the probability of empirically observing this interaction (Catchen et al. 243 2023). Realized interactions occur locally without necessarily being observed (two locally interacting taxa may 244 or may not be seen interacting during sampling), whereas observed interactions are those that have been locally 245 recorded. Local interactions may thus arise from both the ecological (realized interactions) and sampling 246 (observed interactions) processes taking place locally. 247

Local networks are delineated within a particular location and time. We define space as the collection of 248 geographic coordinates (x, y, z), with (x, y) representing longitude and latitude coordinates, and z denoting 249 either altitudes or depths. These point coordinates delineate the spatial boundaries of the system, which may be 250 portrayed as a polyhedron. Ecological interactions may vary along latitudinal and altitudinal gradients, as 251 evidenced by changes in hummingbird-plant interactions (Weinstein & Graham 2017a, b) and mosquito biting 252 rates (e.g., Kulkarni et al. 2006) at different elevations. On the other hand, time is defined as the specific time 253 period within which interactions were either observed or predicted. Even though space and time are continuous 254 variables that should yield probability *densities* of interactions (i.e., relative likelihoods of interactions 255 occurring at infinitesimal locations and instants in time), these definitions enable them to be conceptualized as 256 distinct patches and time segments. Treating space and time as discrete dimensions aligns with the common 257 sampling methods of ecological networks and provides probabilities of interactions, which can be obtained by 258 integrating probability densities over space and time. We can quantify both an area A_0 and a duration t_0 with 259

these definitions. By studying probabilistic local interaction networks, we may thus conduct spatiotemporal
 analyses of local interactions (Box 1), enhancing our understanding of interactions occurring in distinct
 environmental contexts.

²⁶³ What are local probabilistic interactions conditioned on?

²⁶⁴ Local interactions may be conditioned on co-occurrence

The probability that two taxa i and j interact in a local network $L_{x,y,z,t}$ (spatial and temporal subscripts hereafter 265 replaced by the shorter subscript k for clarity) can be conditioned on many environmental and biological factors. 266 In addition to network area (or volume) A_0 and duration t_0 , they may be conditioned on taxa co-occurrence 267 $X_{i,j,k}$, which is usually Boolean, describing if the geographic distributions of both taxa overlap within the study 268 area. As illustrated in Box 1, co-occurrence may be modeled probabilistically, in which case it may conform to 269 a Bernoulli distribution $X_{i,j,k}$ ~ Bernoulli(ϕ), where $\phi = P(X_{i,j,k} = 1)$. The probability of co-occurrence can 270 be calculated using the individual (marginal) occurrence probabilities $P(X_{i,k} = 1)$ and $P(X_{j,k} = 1)$ (which we 271 denote as $P(X_{i,k})$ and $P(X_{j,k})$ for simplicity and better readability). Given that taxa occurrences are not 272 independent of each other, the probability of co-occurrence can be calculated by multiplying the probability of 273 occurrence of one taxon by the probability of occurrence of the other given that the first one is present: 274

$$P(X_{i,j,k}) = P(X_{i,k}, X_{j,k}) = P(X_{i,k} | X_{j,k}) P(X_{j,k}).$$
(1)

²⁷⁵ Note that to keep the text concise and readable, the probability notation used in this manuscript implicitly ²⁷⁶ assigns a value of 1 to binary variables (e.g., in eq. 1 the term $P(X_{i,k}|X_{j,k})$ is short for $P(X_{i,k} = 1|X_{j,k} = 1)$), ²⁷⁷ unless stated otherwise. The value is only stated explicitly when it is 0 or when we wish to emphasize the value ²⁷⁸ of 1.

The probability of co-occurrence $P(X_{i,j,k})$ (short for $P(X_{i,j,k} = 1)$) can be estimated through the application of joint species distribution models (e.g., Pollock *et al.* 2014), potentially taking into account biotic interactions (Staniczenko *et al.* 2017). Given that the probability that two non-co-occurring taxa interact locally is zero (i.e., $P(L_{i,j,k} = 1 | X_{i,j,k} = 0) = 0$), the probability of local interaction can be obtained by multiplying the probability of interaction given co-occurrence with the probability of co-occurrence:

$$P(L_{i,j,k} = 1) = P(L_{i,j,k} = 1 | X_{i,j,k} = 1) \times P(X_{i,j,k} = 1).$$
(2)

²⁸⁴ Knowing that two taxa co-occur improves our estimation of the probability that they interact locally by
 ²⁸⁵ mitigating a potential source of uncertainty.

Local interactions may be conditioned on different environmental and biological factors

Local interactions may also be conditioned on local environmental factors such as temperature (Angilletta et al. 287 2004), precipitation (Woodward et al. 2012), habitat structure (Klecka & Boukal 2014), and the presence or 288 abundance of other taxa in the network (Kéfi et al. 2012; Pilosof et al. 2017). We use the variable E_k to 289 describe the local environmental context in which interaction probabilities were estimated. For example, in a 290 mesocosm experiment estimating interaction probabilities between predators and prey with and without refuges, 291 E_k would represent the presence or absence of these refuges. Like co-occurrence, E_k can also be modeled 292 probabilistically when the variability or uncertainty of environmental factors is considered. E_k represents all 293 environmental variables that were taken into consideration when measuring interaction probabilities; it is 294 therefore a subset of all environmental factors acting on ecological interactions. 295

Other important factors that can impact interaction probabilities at the local scale are taxa local abundances $N_{i,k}$ 296 and $N_{j,k}$, which affect encounter probabilities (Canard *et al.* 2012), and local traits $T_{i,k}$ and $T_{j,k}$ (e.g., movement 297 rates, Beardsell et al. 2021; Cherif et al. 2024), which may also impact encounter probabilities as well as the 298 ability of individuals to interact after encountering each other (Caron et al. 2024; Poisot et al. 2015). Local 299 interaction probabilities may also be conditioned on higher-level properties of the community (i.e., the emerging 300 structure of ecological networks), which we denote by $f(L_k)$. Many topological null models (i.e., statistical 301 models that randomize interactions by retaining certain properties of the network while excluding others) 302 provide interaction probabilities from selected measures of network structure, such as connectance (Fortuna & 303 Bascompte 2006) and the degree distribution (Bascompte et al. 2003). Biological factors, whether at the scale 304 of individual taxa pairs or the community, may thus impact how we estimate and define interaction probabilities. 305

Local interactions may be conditioned on biological feasibility

³⁰⁷ Local interactions must be biologically feasible before occurring at a specific time and space. A local

probability of interaction $P(L_{i,j,k})$ (short for $P(L_{i,j,k} = 1)$) can be expressed as the product of the probability of

local interaction given that the two taxa can potentially interact $P(L_{i,j,k} = 1 | M_{i,j} = 1)$ (which we sometimes denote as $P(L_{i,j,k} | M_{i,j})$ for the sake of simplicity), with their probability of regional interaction $P(M_{i,j} = 1)$:

$$P(L_{i,i,k} = 1) = P(L_{i,i,k} = 1 | M_{i,i} = 1) \times P(M_{i,i} = 1),$$
(3)

assuming that $P(L_{i,j,k} = 1 | M_{i,j} = 0) = 0$.

Low values of $P(L_{i,j,k}|M_{i,j})$ indicate that feasible interactions rarely occur locally, intermediate values around 50% suggest considerable spatiotemporal variability, while high values indicate that regional interactions are nearly always realized locally. The local probability of interaction between a given pair of taxa is thus always equal to or below their probability of regional interaction. Taking into account biological feasibility in our estimation of local interaction probabilities leverages information from the metaweb to better predict the local occurrence of interactions (Dansereau *et al.* 2024; Strydom *et al.* 2021).

318 Conditional variables must be explicitly stated

The probability that two taxa i and j interact in a local network L_k can thus be conditioned on their 319 co-occurrence $X_{i,j,k}$ (or more explicitly on their occurrences $X_{i,k}$ and $X_{j,k}$), local abundances $N_{i,k}$ and $N_{j,k}$, local 320 traits $T_{i,k}$ and $T_{j,k}$, local environmental conditions E_k , network area (or volume) A_0 , time interval t_0 , network 321 properties $f(L_k)$, and biological feasibility $M_{i,j}$. When these conditions are absent from an expression, it may be 322 because they have been marginalized over, which would be reflected in the overall uncertainty of the interaction. 323 Interaction probabilities may also have been implicitly conditioned on missing variables (e.g., when estimated 324 for specific values of these variables without explicitly including them as conditions), potentially impacting our 325 interpretation. The local probability of interaction is described by the following expression when all of these 326 conditional variables are included: 327

$$P(L_{i,j,k}|X_{i,k}, X_{j,k}, N_{i,k}, N_{j,k}, T_{i,k}, T_{j,k}, E_k, A_0, t_0, f(L_k), M_{i,j}).$$

$$\tag{4}$$

These conditional variables do not all need to be considered at all times. The representation of the local context in which probabilities are estimated and the variables that should be taken into consideration depend on the study system, the objectives of the study, and the resources available to the researchers. For example, Gravel *et al.* (2019) analyzed local European host-parasite networks of willow-galling sawflies and their natural enemies,

all referenced in space and time, to infer probabilities of local interactions between co-occurring species. This 332 was achieved by including temperature and precipitation as conditional variables in their models. In Box 2, we 333 reuse these data to show the extent of variation among these local networks. We do so by measuring their 334 dissimilarity with the regional network (metaweb aggregating all local interactions), both in terms of species 335 composition and interactions. We built local probabilistic networks following eq. 3, showing that insufficient 336 local variation (high probability of local interaction among potentially interacting species) results in an 337 overestimation in both the number of interactions and connectance (i.e., the proportion of all of the 338 non-forbidden links that are realized). This analysis was conducted for illustrative purposes, and other 339 conditional variables could have been used to make these comparisons. 340

When accounted for, conditional variables should be clearly described in the documentation of the data 341 (Brimacombe *et al.* 2023), preferentially in mathematical terms to avoid any confusion in their interpretation 342 and to limit manipulation errors during their re-use. For instance, ecologists should be explicit about their 343 consideration $(P(L_{i,j,k}|X_{i,j,k}))$ or not $(P(L_{i,j,k}))$ of co-occurrence in their estimation of local interaction 344 probabilities, as this can change our interpretation of the data and understanding of potential uncertainty 345 sources. Reporting the scale and level of aggregation of the data enables us to more accurately study the 346 underlying ecological processes (Clark et al. 2011) and manipulate or propagate uncertainty to different 347 aggregation levels (Simmonds et al. 2024). In Tbl. 1, we present examples of studies that used different 348 expressions of probabilistic interactions with different conditional variables. We included in this table the 349 probability of empirically observing an interaction that is realized locally $P(O_{i,j,k}|L_{i,j,k})$ to underscore the 350 distinction between local observations and actual realizations of interactions. 351

Table 1: Mathematical expressions of probabilistic interactions. The probability of interaction between two taxa *i* and *j* is interpreted differently in a local network L_k of realized interactions, a local network O_k of observed interactions, a metaweb *M* of potential interactions (representing the *biological* feasibility of interactions), and a metaweb M^* of potential interactions (representing the *ecological* feasibility of interactions). Each expression emphasizes a different conditional variable, the ellipsis serving as a placeholder for other variables not explicitly stated in the expression. The outcome of each of these probabilistic events, along with common models used for estimation, is presented alongside examples of studies that employed them (with specific variables indicated in parentheses, when applicable). The study marked with an asterisk has been conducted on binary interaction networks. The boxes in our study that discuss these expressions are also specified.

Expression	Туре	Outcome	Common models	Reference
$\overline{P(L_{i,j,k} X_{i,k},X_{j,k},\ldots)}$	local	realization of the interaction	species distribution	Gravel et al. (2019),
		given taxa co-occurrence	models	Dansereau et al.
				(2024), Boxes 1 and 5
$P(L_{i,j,k} N_{i,k},N_{j,k},\ldots)$	local	realization of the interaction	neutral models	Canard et al. (2014)
		given taxa abundances		
$P(L_{i,j,k} T_{i,k},T_{j,k},\ldots)$	local	realization of the interaction	trait matching	Caron et al. (2024),
		given local traits	models	Box 4
$P(L_{i,j,k} E_k,\ldots)$	local	realization of the interaction	environmental-	Gravel et al. (2019)
		given local environmental	based models	(temperature and
		conditions		precipitation)
$P(L_{i,j,k} A_0,\ldots)$	local	realization of the interaction in	spatial models	Galiana et al. (2018)
		a given area or volume		*, Box 3
$P(L_{i,j,k} t_0,\ldots)$	local	realization of the interaction	temporal models	Weinstein & Graham
		during a given time period		(2017a), Boxes 1 and
				3
$P(L_{i,j,k} f(L_k),\ldots)$	local	realization of the interaction	topological models	Fortuna & Bascompte
		given network structure		(2006) (connectance),
				Stock et al. (2017)
$P(L_{i,j,k} M_{i,j},)$	local	realization of the interaction	spatiotemporal	Dansereau et al.
		given that it is biologically	models	(2024), Boxes 2, 3,
		feasible		and 5

Expression	Туре	Outcome	Common models	Reference
$\overline{P(O_{i,j,k} L_{i,j,k},\ldots)}$	local	observation of the interaction	sampling models	Catchen et al. (2023)
		given that it is realized locally		
$P(M_{i,j} T_i,T_j)$	regional	biological feasibility of the	trait matching	Strydom et al. (2022),
		interaction given regional	models	Box 4
		traits (non-forbiddenness)		
$P(M^*_{i,j} T_i,T_j,E)$	regional	ecological feasibility of the	trait matching and	this study
		interaction given regional	environmental-	
		traits and environmental	based models	
		conditions		

³⁵² How are local probabilistic interactions estimated?

Various statistical models can be used to estimate local interaction probabilities, some of which are presented in 353 Tbl. 1. These models can be based on multiple conditional variables. Although these variables correspond to 354 distinct ecological inquiries or mechanisms related to ecological interactions, they may covary with each other, 355 such as the possible dependence of $X_{i,j,k}$ and E_k on spatial and temporal scales. When estimating interaction 356 probabilities using e.g. a generalized linear model with multiple explanatory variables that might not all be 357 independent, it may become important to address collinearity. In such cases, it may be necessary to use variable 358 selection techniques before fitting the model to data to mitigate this issue. Other challenges and opportunities 359 associated with predictive models of species interactions are reviewed in Strydom et al. (2021). 360

When using multiple competing models to estimate local interaction probabilities, rather than selecting a single model that best fits the data, model averaging may enhance our estimations. Model weights represent the probability that each model is the most suitable for explaining the data, and may be measured using Akaike weights (Burnham & Anderson 2004; Wagenmakers & Farrell 2004). For instance, given two competing models mod_1 and mod_2 with respective probabilities (or weights) $P(mod_1)$ and $P(mod_2)$, the average probability of interaction $P(L_{i,j,k})$ can be calculated as follows:

$$P(L_{i,j,k}) = P(L_{i,j,k}|mod_1) \times P(mod_1) + P(L_{i,j,k}|mod_2) \times P(mod_2).$$
(5)

Model averaging takes into account the uncertainty of model structure in our estimation of local interaction probabilities. Regardless of the model used for prediction, it is crucial to quantify and disclose all sources of uncertainty to understand better the validity and limitations of our predictions (Simmonds *et al.* 2024).

Box 1: A spatiotemporally explicit model of interactions

Ecologists may resort to predictive models to reconstruct local networks across time and space. We introduce and develop a simple generative Bayesian model for probabilistic local interactions, which explicitly accounts for their spatiotemporal variability. Our model is not designed for regional interactions, which do not vary spatially nor temporally. Rather, it could prove valuable for generating new data on local interactions across time and space, following parameter inference.

As indicated by eq. 2, the probability that two taxa *i* and *j* interact locally can be obtained by multiplying their probability of interaction given co-occurrence with their probability of co-occurrence. The probability of interaction given co-occurrence can be made temporally explicit by modeling it as a Poisson process with rate parameter λ_k . This parameter represents the local expected frequency of interaction between co-occurring taxa. The probability that two co-occurring taxa interact at least once during a time interval t_0 can be given by:

$$P(L_{i,i,k} = 1 | X_{i,i,k} = 1) = 1 - e^{-\lambda_k t_0},$$
(6)

which tends toward 1 as $t_0 \rightarrow \infty$ if $\lambda_k > 0$. In other words, two co-occurring taxa with a nonzero rate of interaction will inevitably interact at least once in a sufficiently long time interval.

The occurrence of an interaction between *i* and *j* may be the result of a Bernoulli trial with parameter ϕ representing the probability of interaction $P(L_{i,j,k} = 1)$. A Bayesian model can be built using the preceding equations to generate new interaction data, following the inference of the λ_k and ϕ parameters.

$$L_{i,j,k} \sim \text{Bernoulli}(\phi)$$
 (7)

$$\phi = P(X_{i,i,k} = 1)(1 - e^{-\lambda_k t_0}) \tag{8}$$

$$P(X_{i,j,k}) \sim \text{Beta}(2,2) \tag{9}$$

$$\lambda_k \sim \text{Exponential}(2)$$
 (10)

In Fig. 1, we show the variation in the probability of interaction under different parameter values. In the right panel, we notice that the probability of interaction always converges toward the probability of cooccurrence $P(X_{i,j,k} = 1)$, for all positive values of the interaction rate.



Figure 1: **Parameters of the spatiotemporally explicit model of interactions.** (a) Probability of local interaction $\phi = P(L_{i,j,k})$ (short for $P(L_{i,j,k} = 1)$) given by the process model (eq. 8) under different values of λ_k (interaction rate) and $P(X_{i,j,k})$ (probability of co-occurrence, short for $P(X_{i,j,k} = 1)$), with $t_0 = 1$ (duration). The probability of local interaction represents the probability that the two taxa will interact at least once within the given time interval. Parameters t_0 and λ_k have complementary units (e.g., t_0 in months and λ_k in number of interactions per month). The parameter values used in the right panel are denoted by the white stars. (b) Scaling of the probability of interaction with the duration parameter t_0 , for different values of λ_k and $P(X_{i,j,k})$.

This model can be customized in different ways, such as linking both parameters to specific environmental or biological variables. For instance, the probability of co-occurrence could be modeled as a function of climatic variables, while the interaction rate parameter could be modeled based on taxa abundances.

Box 2: Dissimilarity of local host-parasite networks

We present a way to assess local network variability and dissimilarity regarding species composition and interactions. We do so by comparing local tripartite host-parasite networks to the metaweb using data from Kopelke *et al.* (2017). This collection of networks consists of interactions between willows, willow-galling sawflies, and their natural enemies sampled across Europe. All data manipulation and methods

³⁷¹

are described in Appendix 1. All code and data to reproduce these analyses are available on Zenodo (https://doi.org/10.5281/zenodo.12802326).

In Fig. 2a-b, we show how the dissimilarity between the metaweb of binary interactions and aggregated local networks changes with the number of sampled local networks. We compared the metaweb and the aggregated local networks using the dissimilarity in species composition (β_S , Fig. 2a) and the dissimilarity of interactions between common species (β_{OS} , Fig. 2b) indices (Poisot *et al.* 2012). Expectedly, local networks are highly dissimilar from the metaweb in terms of species composition, especially when only a limited number of sites have been sampled. This is because few species from the metaweb (species pool) occur locally. Moreover, we observe a peak in the dissimilarity of interactions between common species at intermediate sampling levels. This suggests that species are collected faster than their interactions. With a limited number of sampled local networks, few regional interactions are observed locally. Adding more sites brings new species, but not always their interactions. Quadratic relationships of network properties with sampling effort were also observed by McLeod *et al.* (2021).

373



Figure 2: Network accumulation curves. (a) Dissimilarity in species composition and (b) dissimilarity of interactions between common species between aggregated local networks and the metaweb of binary host-parasite interactions. In both panels, the colored line represents the median dissimilarity across simulations and the grey areas cover the 50% and 95% percentile intervals. (c) Scaling of the number of interactions and (d) scaling of connectance with the number of sampled (aggregated) binary and probabilistic local interaction networks. For a better comparison with binary interactions, local networks of probabilistic interactions were derived from a metaweb of probabilistic interactions with a false positive and false negative rate of zero. A specific value of $P(L_{i,j,k}|M_{i,j})$ (the local probability of interaction among potentially interacting species) was used for all non-aggregated local networks within a particular curve. Aggregated local networks were obtained by sequentially and randomly selecting a number of local networks and aggregating both their species and interactions (with the value of $P(L_{i,j,k}|M_{i,j})$ increasing in aggregated local networks of probabilistic interactions).

Next, we investigate how the number of local interactions and connectance scale with the number of sampled (aggregated) local networks of probabilistic or binary interactions (Fig. 2c-d). By comparing the scaling relationships observed in local networks of binary and probabilistic interactions, we observe that high values of $P(L_{i,j,k}|M_{i,j})$ (short for $P(L_{i,j,k} = 1|M_{i,j} = 1)$) lead to systematic overestimations in the number of interactions and connectance, especially when $P(L_{i,j,k}|M_{i,j}) = 1$ (i.e., when local and

regional probabilities of interactions are equivalent). This suggests that high values of $P(L_{i,j,k}|M_{i,j})$ do not adequately capture the variability of local interactions. However, these biases tend to diminish as the number of sampled networks increases, indicating that most interactions are eventually captured when $P(L_{i,j,k}|M_{i,j})$ is high. In contrast, low values of $P(L_{i,j,k}|M_{i,j})$ lead to missing interactions, resulting in an underestimation of the number of interactions and connectance. These results underscore the importance of using the appropriate level of variability when estimating local interaction probabilities.

375

376 Metawebs: regional catalogs of interactions

377 What are regional probabilistic interactions?

Metawebs (Dunne 2006) are networks of potential interactions over broad spatial, temporal, and taxonomic 378 scales (e.g., food webs at the continental scale). They correspond to the temporal and spatial asymptotes of 379 local interactions (Box 1). Potential interactions describe the biological capacity of taxa to interact under 380 optimal or feasible environmental conditions given enough time, which is typically assessed at the regional 381 scale. Metawebs of probabilistic interactions are particularly useful in situations where there is uncertainty in 382 the ability of taxa to interact (Strydom et al. 2023). They may also be used as informative priors of local 383 interactions. Therefore, building a metaweb of probabilistic interactions may be an important first step before 384 predicting networks at finer scales. 385

In contrast to local networks, where interaction probabilities arise from the variability of interactions and the 386 lack of information on the conditions, interaction probabilities in metawebs solely result from a lack of 387 knowledge. This uncertainty arises due to insufficient interaction data, especially for taxa that have not yet been 388 observed to co-occur, and uncertainties in trait-matching models. As data accumulates, interactions in 389 metawebs should tend towards binarity, either taking a value of 1 (observing an interaction at least once) or 390 approaching 0 (repeatedly failing to observe an interaction between co-occurring taxa). Confidently observing 391 an interaction once confirms its biological feasibility, but failing to observe it (even on multiple occasions) does 392 not ensure that it is non-feasible (e.g., due to false negatives, Catchen et al. 2023). While local interaction 393 probabilities are irreducible because of local variability, the uncertainty of regional interactions reduces to 0 394 with the addition of information. Moreover, although neutrally forbidden interactions (i.e., forbidden 395 interactions between rare species, Canard et al. 2012) have low probability values in local networks, they would 396

have a probability of 1 in the metaweb (this is because the species' traits could support an interaction if they 397 were to encounter each other at high enough abundances). Likewise, non-co-occurring taxa may have a 398 non-zero probability of interaction in the metaweb. Regional interaction probabilities are thus fundamentally 399 different from local interaction probabilities, both in terms of uncertainty sources and probability values. 400 The extent of sampling effort influences our evaluation of probabilities of regional interactions, as sampling 401 over a larger area or for a longer duration enables us to capture a greater number of interactions (Box 1, McLeod 402 et al. 2021). However, in contrast with local networks of probabilistic interactions, regional interactions are not 403 evaluated for any particular local context (they are rather a collection of local contexts), which impacts how they 404 scale with space and time (notably through the extent of the region covered and sampling duration). In Box 3, 405 we discuss the differences in spatial and temporal scaling of regional interactions compared to local

interactions. We do so using the host-parasite networks of Kopelke et al. (2017) as an illustration of spatial 407

scaling (Box 3). Understanding the effect of spatial and temporal scales (including sampling effort) on local and 408

regional interaction probabilities is important for effectively propagating uncertainty across scales and 409

highlighting the fundamental differences between these two types of networks. 410

What are regional probabilistic interactions conditioned on? 41

406

Regional interactions describing biological feasibility are conditioned on traits 412

Potential interactions describe what we refer to as the *biological* feasibility of interactions, which is based solely 413 on the regional traits distributions T_i and T_j of taxa i and j, respectively. We define regional traits distributions 414 as the range of phenotypes that a taxon can express across various environments. Local traits $T_{i,k}$ and $T_{i,k}$, 415 which vary spatially and temporally because of phenotypic plasticity and local environmental variability (Berg 416 & Ellers 2010), are a subset of regional traits. A probability of potential interaction in a metaweb M describing 417 the biological feasibility of interactions may be expressed as: 418

$$P(M_{i,j}|T_i,T_j),\tag{11}$$

which, in contrast with local networks, is not conditioned on any spatial, temporal, co-occurrence or 419 environmental variables (Tbl. 1). Because phylogenetically close species often share similar traits, we should 420 expect that closely related species will have similar interacting partners. We can thus use phylogeny to predict 421

species traits and infer regional interactions (Eklöf & Stouffer 2016; Stouffer *et al.* 2012; Strydom *et al.* 2022).
The taxonomic level at which interactions are evaluated also influences the distribution of regional traits.
However, as explained in Box 4, there is no fundamental difference in the taxonomic scaling of regional and
local interactions (i.e., how interaction probabilities change with taxonomic level) because they both depend on
trait aggregation.

The biological feasibility of interactions expresses our degree of belief that there exists at least one combination of phenotypes that could support a specific type of interaction if they were to encounter each other, assuming they had enough time to interact. Evaluating this probability is conducted without incorporating the environmental conditions under which they encounter each other into the model. It is the complement of the probability $P(F_{i,j}|T_i, T_j)$ of forbidden interactions (i.e., the probability that their traits do not support an interaction), which is based uniquely on biological traits:

$$P(M_{i,j}|T_i, T_j) = 1 - P(F_{i,j}|T_i, T_j).$$
(12)

For example, let *i* be a western diamondback rattlesnake (*Crotalus atrox*) and *j*, a wood lemming (*Myopus*) 433 schisticolor). These two taxa never co-occur, the rattlesnake being adapted to warm regions of North America 434 (Castoe et al. 2007) and the lemming, to northern habitats of Eurasia (Fedorov et al. 2008). As we lack direct 435 observations of an interaction between these two species, we have to rely on expert knowledge or trait-matching 436 models to estimate their probability of potential interaction. To accurately estimate this probability using 437 trait-matching models, it is crucial to ensure that the set of traits considered reflects the overall traits 438 distributions of both taxa. We could for instance consider their average body mass and the average phylogenetic 439 distance of lemmings to rattlesnakes' prey. Doing so, we might find a high probability of potential interaction 440 based on these traits. This example illustrates how regional interactions describing biological feasibility may be 441 estimated solely based on traits, without taking into account environmental conditions (which could be 442 important to consider when e.g. an interaction is forbidden at all temperature values). 443

Regional interactions describing ecological feasibility are conditioned on traits and environmental conditions

The biological feasibility of interactions should not be confused with what we refer to as the *ecological* feasibility of interactions. A probability of potential interaction in a metaweb M^* describing the ecological ⁴⁴⁸ feasibility of interactions may be expressed as:

$$P(M_{i,j}^*|T_i, T_j, E), (13)$$

where E is the environmental conditions under which potential interactions are evaluated (Tbl. 1). Unlike E_k , 449 these environmental conditions do not represent conditions occurring at specific locations. Ecological 450 feasibility represents the probability that two taxa interact if they were to encounter each other under given 451 environmental conditions, assuming they had enough time to interact. Incorporating environmental conditions 452 into a trait-matching model may be important when there is high covariance between the environment and traits. 453 For instance, in our example involving rattlesnakes and lemmings, the probability of potential interaction 454 between these two species may be low in most environmental conditions. Western diamondback rattlesnakes 455 may be unactive under low temperatures (Kissner *et al.* 1997), whereas wood lemmings may have low tolerance 456 to high temperatures (Kausrud et al. 2008). The probability that an interaction is ecologically feasible is always 457 lower than the probability that it is biologically feasible, even across all environmental conditions: 458

$$P(M_{i,j}^*|T_i, T_j) = \int_F P(M_{i,j}^*|T_i, T_j, E)g(E|T_i, T_j)dE \le P(M_{i,j}|T_i, T_j),$$
(14)

where $g(E|T_i, T_j)$ is the conditional probability density function of E given T_i and T_j .

The difference between these two regional probabilities of interaction (across all environmental conditions) arises because the biological feasibility of an interaction is a prerequisite for its ecological feasibility. In other words, biological feasibility is necessary but not sufficient for an interaction to be ecologically feasible. Our discussion of metawebs focuses on the biological feasibility of interactions since most methods developed for inferring probabilities of regional interactions do not explicitly take into account environmental conditions (e.g., Strydom *et al.* 2022).

466 How are regional probabilistic interactions estimated?

467 Starting from a selected set of taxa, which are usually distributed within a broad region of interest, metawebs 468 can be built using different data sources, including literature review (e.g., Maiorano *et al.* 2020), aggregated 469 interaction data (e.g., Gravel *et al.* 2019; Saravia *et al.* 2022), trait-matching models (e.g., Shaw *et al.* 2024; 470 Strydom *et al.* 2022), and expert knowledge, which is not a trivial challenge. Every pair of taxa that have confidently been observed to interact at least once can be given a probability of 1 (i.e., $P(M_{i,j}) = 1$) since we know that they *can* interact. This differs from local networks of probabilistic interactions, where interaction events may remain stochastic (i.e., $P(L_{i,j,k}) < 1$) even after empirically observing interactions due to their spatiotemporal variability. Interactions that were never observed typically have low probability values in local networks and vary from low to high values in metawebs, contingent upon taxa traits distributions (reaching 0 for forbidden links). The aggregation of model predictions and data from different sources thus tends to raise the number of potential interactions in metawebs.

When using local interaction data to estimate probabilities of regional interactions, repeatedly failing to observe an interaction between two co-occurring taxa should decrease the probability that the interaction is biologically feasible. Using Bayes' theorem, the probability that the interaction is biologically feasible given that it was never observed locally, $P(M_{i,j} = 1|O_{i,j,k} = 0)$, may be calculated as follows:

$$P(M_{i,j} = 1 | O_{i,j,k} = 0) = \frac{P(O_{i,j,k} = 0 | M_{i,j} = 1) \times P(M_{i,j} = 1)}{P(O_{i,i,k} = 0)}.$$
(15)

The reduction in the probability of regional interaction after considering that it was never observed locally (i.e., $P(M_{i,j} = 1 | O_{i,j,k} = 0) < P(M_{i,j} = 1)$) occurs because $P(O_{i,j,k} = 0 | M_{i,j} = 1)$ must be lower than $P(O_{i,j,k} = 0)$, i.e. there is a higher chance of observing an interaction when it is biologically feasible.

Observations of interactions may be false positives because of observation errors due to taxonomic 485 misidentifications and ecological misinterpretations, such as those involving phylogenetically close species or 486 cryptic species and interactions (Pringle & Hutchinson 2020). Likewise, forbidden interactions may be false 487 negatives, e.g. if they have been evaluated based on unrepresentative or incomplete traits distributions. 488 Employing Bayesian models proves valuable when estimating interaction probabilities in metawebs (e.g., 489 Bartomeus et al. 2016; Cirtwill et al. 2019). This improvement is achieved by updating prior information 490 regarding the feasibility of interactions (e.g., experts' prior assessments of interaction probabilities) with 491 empirical data on interactions and traits. By improving our estimation of potential interaction probabilities, we 492 may build more reliable metawebs that adequately reflect our uncertainty on the biological feasibility of 493 interactions. 494

Box 3: Spatial and temporal scaling of interactions

Local networks and metawebs have distinct relationships with space (area or volume) and time (sampling effort or duration). Local probabilities of interaction scale both spatially and temporally, because local interactions have more opportunities to be realized in larger areas and longer durations. In a larger sampling area and duration, we increase the likelihood of sampling favorable conditions for interactions to occur. If a local network of probabilistic interactions L_1 with an area A_1 is compared to a larger network L_0 with an area A_0 , and A_1 is entirely nested within A_0 , interaction probabilities should be lower in the smaller network, i.e. $P(L_{i,j,1}|A_1 < A_0) \leq P(L_{i,j,0}|A_0)$. However, if A_1 and A_0 are disjoint, interaction probabilities could be higher in the smaller area, contingent upon local environmental and biological conditions. In contrast, regional probabilities of interaction do not scale with space and time. The probability of two taxa potentially interacting should be the same in all metawebs in which they are present regardless of scale, provided that the data and methods used for estimation are consistent. This is because they depend solely on the biological capacity of two taxa to interact, regardless of co-occurrence and local environmental conditions. However, probabilities of regional interactions may change, tending to become more definitive, with increased sampling effort.

In Fig. 3, we show how the expected *number* of local host-parasite interactions scales with the spatial boundary of the network (represented by an expanding latitudinal window) in comparison with regional interactions. We do so using the host-parasite networks of Kopelke *et al.* (2017). The increase in the number of regional interactions is due to the inclusion of more species in a larger area. To ensure a conservative comparison between aggregated local and regional networks, we employed equal interaction probabilities (i.e., using $P(L_{i,j,k}|M_{i,j}) = 1$) in both types of network. This means that local interaction probabilities could not increase further when aggregating them. Despite this, we notice that the total number of regional interactions scales more rapidly than local interactions. This is because numerous regional interactions involve species that never co-occur, and as a result, these interactions are not captured in local networks. All data manipulation and methods are described in Appendix 1.

495



Figure 3: **Spatial scaling of interactions.** Expected number of host-parasite interactions in a network aggregating all (a) local and (b) regional probabilistic interactions within a latitudinal window of a given width. Every dashed curve corresponds to a different window centered at a given latitude (color bar), with the pink solid line representing the median number of interactions across windows. Heatmaps of the expected number of (c) local and (d) regional interactions found in windows of specified width and position (central latitude). Probabilities of regional interactions were obtained with a false positive rate of 5% and a false negative rate of 10%. Local probabilistic interactions were derived from regional probabilistic interactions by setting the value of $P(L_{i,j,k}|M_{i,j})$ (the local probability of interaction among potentially interactions found within a particular latitudinal window, with the values of $P(L_{i,j,k}|M_{i,j})$ remaining at their maximum value of 1.

496

Box 4: Taxonomic scaling of interactions

Given that our interpretation of the properties of ecological networks depends on their taxonomic level

(Melián et al. 2011), investigating the taxonomic scaling of interactions (i.e., how interaction probabilities

change with taxonomic level) is important. There are no inherent differences between the taxonomic scaling of local and regional interactions. The taxonomic level of interactions impacts the definition of nodes. Local and regional interaction probabilities are not directly conditioned on taxonomic scale. However, some conditional variables (e.g., trait distribution) may covary with taxonomic scale. In such cases, local and regional interaction probabilities would change taxonomically following the scaling of these variables.

In both types of interactions, transitioning to a broader level of organization (e.g., from a species-level network *S* to a genus-level network *G*) can be done using interaction probabilities from finer scales. For example, in a network with n_1 species of genus g_1 and n_2 species of genus g_2 , one can calculate the probability that at least one species from genus g_1 interacts with at least one species from genus g_2 (i.e., the probability that the genus-level interaction occurs) as follows:

$$P(G_{g_1,g_2}) = 1 - \prod_{i=1}^{n_1} \prod_{j=1}^{n_2} (1 - P(S_{g_{1,i},g_{2,j}})),$$
(16)

where $g_{1,i}$ and $g_{2,j}$ are the species of the corresponding genus. This equation assumes independence between species-level interactions, which may not hold true in practice due to the strong phylogenetic signal frequently encountered in species interactions (Gomez *et al.* 2010). In contrast, a different approach is necessary when transitioning from a broader to a finer level of organization. This is because the knowledge of an interaction between two genera does not guarantee that all possible pairwise species combinations will also interact. One possible method is to build a finer-scale network by generating probabilities of interaction through random sampling from a beta distribution, parameterized by the broader-scale network.

Fundamentally, the taxonomic scaling of interactions involves aggregating interactions between individuals into larger groups. Interaction probabilities at broader taxonomic scales should thus conform to probabilities of interactions between individuals. For example, Canard *et al.* (2012) built a species-based network using simulated individual-based networks. In local individual-based food webs, the probability that two individuals interact reflects our degree of belief that one individual will consume the other. Likewise, in local species-based food webs, the probability that two species interact represents our degree of belief that *at least* one individual from the predator species will consume at least another individual from the prey species. In that regard, taxonomic scaling is analogous to the spatial and temporal scaling

of interactions, as they all represent different ways to aggregate individuals into broader groups (either spatially, temporally, or taxonomically).

Box 5: Sampling for binary interaction networks

Local networks of binary interactions may be predicted by performing independent Bernoulli trials for each probabilistic interaction. This is particularly useful when analyzing the structure of probabilistic interaction networks in the absence of specific analytical formulas (Poisot *et al.* 2016), even though it may introduce biases in our estimations when connectance is low (Chagnon 2015; Poisot & Gravel 2014). There are at least two techniques to sampling binary interaction networks across space, each predicting a binary interaction network for each location k within a given region. The first technique involves performing a single Bernoulli trial for each pair of taxa based on their regional probability of interaction:

$$M_{i,j} \sim \text{Bernoulli}(\phi_{i,j}),$$

where $\phi_{i,j} = P(M_{i,j} = 1)$.

In employing this technique, we predict a single metaweb of binary interactions for each simulation. Every pair of taxa predicted to interact in this metaweb will be treated as interacting in all localized networks where they co-occur, i.e. $L_{i,j,k} = M_{i,j}$ when $X_{i,j,k} = 1$. This will result in local pairwise interactions without spatial variation.

The second technique is to independently sample each local network of probabilistic interactions:

$$L_{i,j,k} \sim \text{Bernoulli}(\phi_{i,j,k}),$$

where $\phi_{i,j,k} = P(L_{i,j,k} = 1)$.

500

This can be achieved by first generating distinct probabilistic interaction networks for each location. Because binary interactions are sampled independently for each location, this second technique captures network structure across space and time more effectively. When sampling binary interactions from local interaction probabilities, it is crucial to sample at the same spatial scale for which probabilities were estimated to prevent systematic biases in predictions.

In Fig. 4, we compare the average connectance of binary interaction networks resulting from these two

sampling techniques. We sampled regional and local interactions from our host-parasite networks of probabilistic interactions (Kopelke *et al.* 2017), generating a number of binary interaction network realizations for each site in the dataset. These two sampling techniques yield different outcomes, particularly for intermediate values of $P(L_{i,j,k}|M_{i,j})$ of 0.50, which represent instances where regional interactions do not consistently manifest locally (i.e., with the largest local variability). As anticipated, we observe that sampling binary interactions from the metaweb tends to overestimate connectance on average compared to sampling them from local networks (Fig. 4). We also observe an increase in the variability of connectance when employing a single simulation (Fig. 4a-c, cross markers), which is a more tangible representation of the process leading to the realization of local interactions in nature. All data manipulation and methods are described in Appendix 1.



Figure 4: **Connectance of sampled binary interaction networks.** (a-c) Average connectance of binary interaction networks obtained from the two sampling techniques for 20 randomly selected hostparasite networks. Cross markers represent the connectance of a single sample for each network, diamond markers the average connectance across 10 samples, hexagon markers the average connectance across 50 samples, and the colored circles the average connectance across 100 samples (marker size proportional to the number of samples). (d-f) Reduction in the mean squared logarithmic error between the average connectance of binary interaction networks (all 233 host-parasite networks) obtained from these two sampling techniques as the number of samples increases. The local probability of interaction between potentially interacting species was set to three different values: (a,d) $P(L_{i,j,k}|M_{i,j}) = 1.0$, (b,e) $P(L_{i,j,k}|M_{i,j}) = 0.75$, and (c,f) $P(L_{i,j,k}|M_{i,j}) = 0.50$. Probabilities of regional interactions were obtained with a false positive rate of 5% and a false negative rate of 10%. Regional samples were obtained by randomly sampling binary interactions from the species potentially engaged in the interactions. Local samples were obtained by independently sampling binary interactions for each local network of probabilistic interactions.

Both sampling techniques assume independence between interactions, which might not hold true in reality. Covariation among interactions could exist even if we do not explicitly condition interactions on others. For example, the probability that two taxa interact could change with the realization of another interaction or the presence or abundance of other taxa (Kéfi *et al.* 2012; Pilosof *et al.* 2017). The consequences of this assumption of independence on the prediction of network structure have yet to be empirically examined. Sampling whole networks (or graphs) instead of pairwise interactions may eliminate the need for this assumption of independence (Battiston *et al.* 2020).

Future perspectives

In this contribution, we underline the importance of network documentation for adequately interpreting and 504 manipulating probabilistic interaction data. The mathematical representation of probabilities and their statistical 505 properties depend on the type of interactions (local or regional) and the conditions under which these 506 interactions were evaluated. We show that local networks and metawebs of probabilistic interactions differ in 507 their relationship to spatial and temporal scales (Box 3), with regional interactions remaining consistent across 508 scales. In contrast with metawebs, local interactions are measured in a specific context (e.g., in a given area, 509 time, and biological and environmental conditions) and depend on taxa co-occurrence. These differences bring 510 to light the need to use probabilistic data with caution, for instance when generating network realizations of 511 binary interactions across space (Box 5). Clear documentation describing the type of interaction and the 512 variables used in their estimation are required to ensure adequate data manipulation. Sound data practices and 513 foundations for probabilistic thinking in network ecology facilitate reliable assessments of the spatiotemporal 514 variability and uncertainty of biotic interactions. Here we identify key research priorities for improving our 515 understanding of probabilistic local and regional interactions. 516

517 Predicting local networks from metawebs

Metawebs are a valuable source of ecological information for predicting local networks across time and space. 518 Local networks of binary interactions can be reconstructed by selecting a subset of taxa and interactions from 519 the metaweb (Dunne 2006). Determining the list of taxa to select can be achieved empirically (e.g., observed 520 occurrence data for a site) or numerically (e.g., species distribution models). As species composition is 521 arguably easier to sample and predict than pairwise interactions, the primary challenge lies in deciding which 522 interactions to select from the metaweb. Inferring the structure of local networks from the metaweb before 523 predicting local pairwise interactions could hold promise (Strydom et al. 2021), considering that the structure 524 of local networks is constrained by the metaweb (Saravia et al. 2022). 525

While predicting local binary interactions from a metaweb is not be a simple task, inferring local networks of probabilistic interactions from a metaweb comes with its own set of challenges. For example, Dansereau *et al.* (2024) inferred spatially-explicit food webs from a metaweb of probabilistic trophic interactions between Canadian mammals. Their predicted localized food webs are downscaled versions of the metaweb (i.e., localized metawebs with the same interaction probabilities as those in the regional metaweb). To infer local networks as defined in this manuscript (i.e., describing local realizations of interactions), local interaction
 probabilities must be smaller than regional interaction probabilities. Inferring local networks from a metaweb
 by maintaining identical interaction probability values introduces systematic biases into the predictions, as
 discussed in Box 2 (unless networks are seen as downscaled metawebs).

As suggested by McLeod *et al.* (2021), metawebs establish an upper limit for local interactions (similarly for metawebs of probabilistic interactions, Strydom *et al.* 2023). In other words, the probability that two taxa interact at a specific location and time is consistently lower or equal to the probability of their regional interaction, regardless of the conditional variables considered:

$$P(L_{i,j,k}|...) \le P(M_{i,j}|T_i, T_j).$$
(17)

Moreover, the probability that two taxa possess the biological capacity to interact must be higher than the
probability of them interacting at any location and time because they may never co-occur or encounter locally.
Specifically, the marginal probability of local interaction across all spatial, temporal, and environmental
conditions must be less than the probability of regional interaction, i.e.

$$\int_{E_k} \int_{A_0} \int_{t_0} P(L_{i,j,k} | E_k, A_0, t_0) g(E_k, A_0, t_0) \, \mathrm{d}t_0 \, \mathrm{d}A_0 \, \mathrm{d}E_k \le P(M_{i,j} | T_i, T_j), \tag{18}$$

where $g(E_k, A_0, t_0)$ is the joint density function of E_k, A_0 , and t_0 .

Estimating more precisely the probability $P(L_{i,j,k}|M_{i,j})$ that two taxa interact locally if they can potentially interact allows for improved predictions of local networks from the metaweb of probabilistic interactions. This task is challenging due to the variability of this probability across space and time, as well as its variability across pairwise interactions within a network. Using simple models of $P(L_{i,j,k}|M_{i,j})$, as shown in Appendix 1, represents an initial step toward the overarching objective of reconstructing local networks from metawebs.

549 Quantifying and reducing interaction uncertainty

⁵⁵⁰ While sampling biological communities decreases the uncertainty of interactions by accumulating evidence for

their feasibility and local realization, there is a limit to how much we can reduce uncertainty. In metawebs,

probabilities reflect our limited knowledge of interactions, which is expected to improve with a larger volume of

data. Regional interactions should become more definitive (with probabilities approaching 0 or 1) as we investigate various conditions, including different combinations of species traits.

In comparison, local interaction probabilities represent both our knowledge uncertainty and their spatiotemporal 555 variability. Owing to environmental heterogeneity, there will invariably be instances in which an interaction 556 occurs and others in which it does not, across different times and locations, irrespective of the extent to which 557 we can improve our knowledge of its biological feasibility and the local conditions that facilitate its occurrence. 558 When local networks describe probabilities of observing interactions rather than their realization, we must also 559 consider observation uncertainty (sampling error) as an additional source of uncertainty. Quantifying and 560 partitioning this uncertainty will enable us to make more accurate predictions about ecological interactions at 561 various spatial and temporal scales, and to identify priority sampling locations to reduce this uncertainty. This 562 will prove to be of vital importance as our time to understand nature runs out, especially at locations where the 563 impacts of climate change and habitat loss hit harder. 564

Relaxing the independence assumption

Estimating local interaction probabilities independently for each taxa pair and assembling them into a network 566 of probabilistic interactions comes with limitations. Predicting local networks of binary interactions based on 567 these interaction probabilities assumes independence among interactions, a condition seldom respected in 568 practice (Golubski & Abrams 2011). The occurrence of an interaction may depend on the realization of other 569 interactions or the presence or abundance of other taxa in the network (Kéfi et al. 2012; Pilosof et al. 2017). 570 Relaxing this assumption of independence is the next logical step in the stochastic representation of interactions. 571 A more accurate representation of the uncertainty and variability of ecological networks involves creating 572 probabilistic networks ($P(L_k)$ and P(M)), rather than networks of probabilistic interactions ($P(L_{i,i,k})$) and 573 $P(M_{i})$). Probabilistic networks describe the probability that a particular network of binary (or quantitative) 574 interactions (its whole adjacency matrix) is realized. For example, Young et al. (2021) used a Bayesian 575 approach to estimate the probability of different plant-pollinator network structures derived from imperfect 576 observational data. A probability distribution of ecological networks may also be derived using the principle of 577 maximum entropy given structural constrained (e.g., Cimini et al. 2019; Park & Newman 2004). 578 Regardless of the method used, generating probabilistic local networks could lead to more accurate predictions 579 of local networks of binary interactions by bypassing the independence assumption. Probabilistic networks 580

could serve as an alternative to null hypothesis significance testing when comparing the structure of a local 581 network to some random expectations or, as done in Pellissier et al. (2018) and Box 2, to the metaweb. These 582 random expectations are typically derived by performing a series of Bernoulli trials on probabilistic 583 interactions, assuming independence, to generate a distribution of networks of binary interactions to calculate 584 their structure (Poisot *et al.* 2016). One could instead compare the likelihood of an observed network to the one 585 of the most likely network structure (according to the probabilistic network distribution), thereby directly 586 obtaining a measure of discrepancy of the empirical network. Generating probabilistic ecological networks 587 represents a tangible challenge, one that, in the coming years, promises to unlock doors to more advanced and 588 adequate analyses of ecological networks. 589

590 Acknowledgment

⁵⁹¹ We acknowledge that this study was conducted on land within the traditional unceded territory of the Saint
⁵⁹² Lawrence Iroquoian, Anishinabewaki, Mohawk, Huron-Wendat, and Omàmiwininiwak nations. This work was
⁵⁹³ supported by the Institute for Data Valorisation (IVADO) and the Natural Sciences and Engineering Research
⁵⁹⁴ Council of Canada (NSERC) Collaborative Research and Training Experience (CREATE) program, through the
⁵⁹⁵ Computational Biodiversity Science and Services (BIOS²) program. A special thanks to all members of the
⁵⁹⁶ Black Holes and Revelations working group (organized by BIOS²) and the Poisot Lab for their insightful
⁵⁹⁷ discussions and valuable feedback on this manuscript.

598 **References**

- Abrego, N., Roslin, T., Huotari, T., Ji, Y., Schmidt, N.M., Wang, J., et al. (2021). Accounting for species
- interactions is necessary for predicting how arctic arthropod communities respond to climate change.
- 601 *Ecography*, 44, 885–896.
- Angilletta, M.J., Jr., Steury, T.D. & Sears, M.W. (2004). Temperature, Growth Rate, and Body Size in
- Ectotherms: Fitting Pieces of a Life-History Puzzle1. *Integrative and Comparative Biology*, 44, 498–509.
- Araujo, M.B., Rozenfeld, A., Rahbek, C. & Marquet, P.A. (2011). Using species co-occurrence networks to
- assess the impacts of climate change. *Ecography*, 34, 897–908.

- Bartomeus, I., Gravel, D., Tylianakis, J.M., Aizen, M.A., Dickie, I.A. & Bernard-Verdier, M. (2016). A
 common framework for identifying linkage rules across different types of interactions. *Functional Ecology*,
 30, 1894–1903.
- Bascompte, J., Jordano, P., Melián, C.J. & Olesen, J.M. (2003). The nested assembly of plant–animal
 mutualistic networks. *Proceedings of the National Academy of Sciences*, 100, 9383–9387.
- Battiston, F., Cencetti, G., Iacopini, I., Latora, V., Lucas, M., Patania, A., et al. (2020). Networks beyond
- pairwise interactions: Structure and dynamics. *Physics Reports*, Networks beyond pairwise interactions:
- ⁶¹³ Structure and dynamics, 874, 1–92.
- ⁶¹⁴ Beardsell, A., Gravel, D., Berteaux, D., Gauthier, G., Clermont, J., Careau, V., et al. (2021). Derivation of

Predator Functional Responses Using a Mechanistic Approach in a Natural System. *Frontiers in Ecology*

616 and Evolution, 9.

- Berg, M.P. & Ellers, J. (2010). Trait plasticity in species interactions: A driving force of community dynamics.
 Evolutionary Ecology, 24, 617–629.
- ⁶¹⁹ Berlow, E.L., Neutel, A.-M., Cohen, J.E., De Ruiter, P.C., Ebenman, B., Emmerson, M., *et al.* (2004).

Interaction strengths in food webs: Issues and opportunities. *Journal of Animal Ecology*, 73, 585–598.

- Bolnick, D.I., Amarasekare, P., Araújo, M.S., Bürger, R., Levine, J.M., Novak, M., et al. (2011). Why
- intraspecific trait variation matters in community ecology. Trends in Ecology & Evolution, 26, 183–192.
- Borrett, S.R. & Scharler, U.M. (2019). Walk partitions of flow in Ecological Network Analysis: Review and
- synthesis of methods and indicators. *Ecological Indicators*, 106, 105451.
- Boulangeat, I., Gravel, D. & Thuiller, W. (2012). Accounting for dispersal and biotic interactions to disentangle
 the drivers of species distributions and their abundances. *Ecology Letters*, 15, 584–593.
- ⁶²⁷ Brimacombe, C., Bodner, K. & Fortin, M.-J. (2024). Applying a method before its proof of concept: A
- cautionary tale using inferred food webs. *Global Change Biology*, 30, e17360.
- Brimacombe, C., Bodner, K., Michalska-Smith, M., Poisot, T. & Fortin, M.-J. (2023). Shortcomings of reusing
 species interaction networks created by different sets of researchers. *PLOS Biology*, 21, e3002068.
- Broom, M. & Ruxton, G.D. (2005). You can run—or you can hide: Optimal strategies for cryptic prey against
- ⁶³² pursuit predators. *Behavioral Ecology*, 16, 534–540.

- Burnham, K.P. & Anderson, D.R. (2004). Multimodel Inference: Understanding AIC and BIC in Model
- ⁶³⁴ Selection. Sociological Methods & Research, 33, 261–304.
- ⁶³⁵ Canard, E.F., Mouquet, N., Mouillot, D., Stanko, M., Miklisova, D. & Gravel, D. (2014). Empirical Evaluation
 ⁶³⁶ of Neutral Interactions in Host-Parasite Networks. *The American Naturalist*, 183, 468–479.
- ⁶³⁷ Canard, E., Mouquet, N., Marescot, L., Gaston, K.J., Gravel, D. & Mouillot, D. (2012). Emergence of
- 638 Structural Patterns in Neutral Trophic Networks. *PLOS ONE*, 7, e38295.
- ⁶³⁹ Carlson, C.J., Farrell, M.J., Grange, Z., Han, B.A., Mollentze, N., Phelan, A.L., *et al.* (2021). The future of
 ⁶⁴⁰ zoonotic risk prediction. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 376,
 ⁶⁴¹ 20200358.
- ⁶⁴² Caron, D., Brose, U., Lurgi, M., Blanchet, F.G., Gravel, D. & Pollock, L.J. (2024). Trait-matching models
- predict pairwise interactions across regions, not food web properties. *Global Ecology and Biogeography*,
 33, e13807.
- ⁶⁴⁵ Castoe, T.A., Spencer, C.L. & Parkinson, C.L. (2007). Phylogeographic structure and historical demography of
 ⁶⁴⁶ the western diamondback rattlesnake (*Crotalus Atrox*): A perspective on North American desert

biogeography. *Molecular Phylogenetics and Evolution*, 42, 193–212.

- Catchen, M.D., Poisot, T., Pollock, L.J. & Gonzalez, A. (2023). The missing link: Discerning true from false
 negatives when sampling species interaction networks.
- ⁶⁵⁰ Cazelles, K., Araujo, M.B., Mouquet, N. & Gravel, D. (2016). A theory for species co-occurrence in interaction
 ⁶⁵¹ networks. *Theoretical Ecology*, 9, 39–48.
- ⁶⁵² Chagnon, P.-L. (2015). Characterizing topology of ecological networks along gradients: The limits of metrics'
 ⁶⁵³ standardization. *Ecological Complexity*, 22, 36–39.
- ⁶⁵⁴ Cherif, M., Brose, U., Hirt, M.R., Ryser, R., Silve, V., Albert, G., *et al.* (2024). The environment to the rescue:
 ⁶⁵⁵ Can physics help predict predator-prey interactions? *BIOLOGICAL REVIEWS*.
- ⁶⁵⁶ Choh, Y., Ignacio, M., Sabelis, M.W. & Janssen, A. (2012). Predator-prey role reversals, juvenile experience
 ⁶⁵⁷ and adult antipredator behaviour. *Scientific Reports*, 2, 728.
- ⁶⁵⁸ Cimini, G., Squartini, T., Saracco, F., Garlaschelli, D., Gabrielli, A. & Caldarelli, G. (2019). The statistical
- ⁶⁵⁹ physics of real-world networks. *Nature Reviews Physics*, 1, 58–71.

- ⁶⁶⁰ Cirtwill, A.R., Eklöf, A., Roslin, T., Wootton, K. & Gravel, D. (2019). A quantitative framework for
- investigating the reliability of empirical network construction. *Methods in Ecology and Evolution*, 10,
 902–911.
- ⁶⁶³ Clark, J.S., Bell, D.M., Hersh, M.H., Kwit, M.C., Moran, E., Salk, C., *et al.* (2011). Individual-scale variation,
 ⁶⁶⁴ species-scale differences: Inference needed to understand diversity. *Ecology Letters*, 14, 1273–1287.
- Dallas, T., Park, A.W. & Drake, J.M. (2017). Predicting cryptic links in host-parasite networks. *PLOS Computational Biology*, 13, e1005557.
- Dansereau, G., Barros, C. & Poisot, T. (2024). Spatially explicit predictions of food web structure from
 regional-level data. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 379, 20230166.
- ⁶⁶⁹ Dunne, J.A. (2006). The Network Structure of Food Webs. In: *Ecological networks: Linking structure and*
- *dynamics* (eds. Dunne, J.A. & Pascual, M.). Oxford University Press, pp. 27–86.
- ⁶⁷¹ Eklöf, A. & Stouffer, D.B. (2016). The phylogenetic component of food web structure and intervality.
- ⁶⁷² *Theoretical Ecology*, 9, 107–115.
- Elmasri, M., Farrell, M.J., Davies, T.J. & Stephens, D.A. (2020). A hierarchical Bayesian model for predicting
- ecological interactions using scaled evolutionary relationships. *The Annals of Applied Statistics*, 14,
- ⁶⁷⁵ 221–240.
- Elton, C.S. (2001). Animal Ecology. University of Chicago Press, Chicago, IL.
- Emmerson, M.C. & Raffaelli, D. (2004). Predator-prey body size, interaction strength and the stability of a real
- food web. Journal of Animal Ecology, 73, 399–409.
- ⁶⁷⁹ Fedorov, V.B., Goropashnaya, A.V., Boeskorov, G.G. & Cook, J.A. (2008). Comparative phylogeography and
- demographic history of the wood lemming (Myopus schisticolor): Implications for late Quaternary history
- of the taiga species in Eurasia. *Molecular Ecology*, 17, 598–610.
- Fortuna, M.A. & Bascompte, J. (2006). Habitat loss and the structure of plant–animal mutualistic networks.
 Ecology Letters, 9, 281–286.
- ⁶⁸⁴ Fu, X., Seo, E., Clarke, J. & Hutchinson, R.A. (2021). Link Prediction Under Imperfect Detection:
- 685 Collaborative Filtering for Ecological Networks. *IEEE Transactions on Knowledge and Data Engineering*,
- ⁶⁸⁶ 33, 3117–3128.

- Galiana, N., Lurgi, M., Claramunt-López, B., Fortin, M.-J., Leroux, S., Cazelles, K., *et al.* (2018). The spatial
 scaling of species interaction networks. *Nature Ecology & Evolution*, 2, 782–790.
- Gilman, S.E., Urban, M.C., Tewksbury, J., Gilchrist, G.W. & Holt, R.D. (2010). A framework for community
 interactions under climate change. *Trends in Ecology & Evolution*, 25, 325–331.
- Godsoe, W., Murray, R. & Iritani, R. (2022). Species interactions and diversity: A unified framework using Hill
 numbers. *Oikos*, n/a, e09282.
- ⁶⁹³ Golubski, A.J. & Abrams, P.A. (2011). Modifying modifiers: What happens when interspecific interactions
 ⁶⁹⁴ interact? *Journal of Animal Ecology*, 80, 1097–1108.
- Gomez, J.M., Verdu, M. & Perfectti, F. (2010). Ecological interactions are evolutionarily conserved across the
 entire tree of life. *Nature*, 465, 918–U6.
- Gonzalez, A. & Londoño, M.C. (2022). Monitor biodiversity for action. Science, 378, 1147–1147.
- Gonzalez-Varo, J.P. & Traveset, A. (2016). The Labile Limits of Forbidden Interactions. *Trends in Ecology & Evolution*, 31, 700–710.
- Grabowski, J.H. (2004). Habitat Complexity Disrupts Predator–Prey Interactions but Not the Trophic Cascade
 on Oyster Reefs. *Ecology*, 85, 995–1004.
- ⁷⁰² Gravel, D., Baiser, B., Dunne, J.A., Kopelke, J.-P., Martinez, N.D., Nyman, T., et al. (2019). Bringing Elton
- and Grinnell together: A quantitative framework to represent the biogeography of ecological interaction
- ⁷⁰⁴ networks. *Ecography*, 42, 401–415.
- Gravel, D., Poisot, T., Albouy, C., Velez, L. & Mouillot, D. (2013). Inferring food web structure from
 predator–prey body size relationships. *Methods in Ecology and Evolution*, 4, 1083–1090.
- ⁷⁰⁷ Guimarães, P.R. (2020). The Structure of Ecological Networks Across Levels of Organization. *Annual Review* ⁷⁰⁸ of Ecology, Evolution, and Systematics, 51, 433–460.
- ⁷⁰⁹ Guimerà, R. & Sales-Pardo, M. (2009). Missing and spurious interactions and the reconstruction of complex
 ⁷¹⁰ networks. *Proceedings of the National Academy of Sciences*, 106, 22073–22078.
- Hemprich-Bennett, D.R., Oliveira, H.F.M., Le Comber, S.C., Rossiter, S.J. & Clare, E.L. (2021). Assessing the
- ⁷¹² impact of taxon resolution on network structure. *Ecology*, 102, e03256.

- Herrera, C.M. (1989). Pollinator abundance, morphology, and flower visitation rate: Analysis of the "quantity"
 component in a plant-pollinator system. *Oecologia*, 80, 241–248.
- Hortal, J., de Bello, F., Diniz-Filho, J.A.F., Lewinsohn, T.M., Lobo, J.M. & Ladle, R.J. (2015). Seven Shortfalls
- that Beset Large-Scale Knowledge of Biodiversity. Annual Review of Ecology, Evolution, and Systematics,
 46, 523–549.
- Jordano, P. (1987). Patterns of Mutualistic Interactions in Pollination and Seed Dispersal: Connectance,
- ⁷¹⁹ Dependence Asymmetries, and Coevolution. *The American Naturalist*, 129, 657–677.
- Jordano, P. (2016). Sampling networks of ecological interactions. *Functional Ecology*, 30, 1883–1893.

Jordano, P., Bascompte, J. & Olesen, J.M. (2003). Invariant properties in coevolutionary networks of
 plant–animal interactions. *Ecology Letters*, 6, 69–81.

- Kausrud, K.L., Mysterud, A., Steen, H., Vik, J.O., Østbye, E., Cazelles, B., *et al.* (2008). Linking climate
 change to lemming cycles. *Nature*, 456, 93–97.
- Kéfi, S., Berlow, E.L., Wieters, E.A., Joppa, L.N., Wood, S.A., Brose, U., *et al.* (2015). Network structure
 beyond food webs: Mapping non-trophic and trophic interactions on Chilean rocky shores. *Ecology*, 96, 291–303.
- Kéfi, S., Berlow, E.L., Wieters, E.A., Navarrete, S.A., Petchey, O.L., Wood, S.A., *et al.* (2012). More than a
 meal... integrating non-feeding interactions into food webs. *Ecology Letters*, 15, 291–300.
- ⁷³⁰ Kéfi, S., Miele, V., Wieters, E.A., Navarrete, S.A. & Berlow, E.L. (2016). How Structured Is the Entangled
- Bank? The Surprisingly Simple Organization of Multiplex Ecological Networks Leads to Increased
- Persistence and Resilience. *PLOS Biology*, 14, e1002527.
- ⁷³³ Kissner, K.J., Forbes, M.R. & Secoy, D.M. (1997). Rattling Behavior of Prairie Rattlesnakes (Crotalus viridis

viridis, Viperidae) in Relation to Sex, Reproductive Status, Body Size, and Body Temperature. *Ethology*,

- 103, 1042–1050.
- ⁷³⁶ Klecka, J. & Boukal, D.S. (2014). The effect of habitat structure on prey mortality depends on predator and
 ⁷³⁷ prey microhabitat use. *Oecologia*, 176, 183–191.
- ⁷³⁸ Kopelke, J.-P., Nyman, T., Cazelles, K., Gravel, D., Vissault, S. & Roslin, T. (2017). Food-web structure of
- willow-galling sawflies and their natural enemies across Europe. *Ecology*, 98, 1730–1730.

- ⁷⁴⁰ Kulkarni, M.A., Kweka, E., Nyale, E., Lyatuu, E., Mosha, F.W., Chandramohan, D., et al. (2006).
- Entomological Evaluation of Malaria Vectors at Different Altitudes in Hai District, Northeastern Tanzania.
- Journal of Medical Entomology, 43, 580–588.
- ⁷⁴³ Maiorano, L., Montemaggiori, A., Ficetola, G.F., O'Connor, L. & Thuiller, W. (2020). TETRA-EU 1.0: A
- species-level trophic metaweb of European tetrapods. *Global Ecology and Biogeography*, 29, 1452–1457.
- ⁷⁴⁵ McCann, K. (2007). Protecting biostructure. *Nature*, 446, 29–29.
- 746 McCann, K.S. (2011). Food webs (MPB-50). In: Food Webs (MPB-50). Princeton University Press.
- McLeod, A., Leroux, S.J., Gravel, D., Chu, C., Cirtwill, A.R., Fortin, M.-J., *et al.* (2021). Sampling and
 asymptotic network properties of spatial multi-trophic networks. *Oikos*, n/a.
- ⁷⁴⁹ Melián, C.J., Vilas, C., Baldó, F., González-Ortegón, E., Drake, P. & Williams, R.J. (2011). Eco-evolutionary
- ⁷⁵⁰ Dynamics of Individual-Based Food Webs. In: *Advances in Ecological Research*, The Role of Body Size in
- ⁷⁵¹ Multispecies Systems (ed. Belgrano, A.). Academic Press, pp. 225–268.
- ⁷⁵² Morales-Castilla, I., Matias, M.G., Gravel, D. & Araújo, M.B. (2015). Inferring biotic interactions from
 ⁷⁵³ proxies. *Trends in Ecology & Evolution*, 30, 347–356.
- ⁷⁵⁴ Ohlmann, M., Miele, V., Dray, S., Chalmandrier, L., O'Connor, L. & Thuiller, W. (2019). Diversity indices for
- recological networks: A unifying framework using Hill numbers. *Ecology Letters*, 22, 737–747.
- Olesen, J.M., Bascompte, J., Dupont, Y.L., Elberling, H., Rasmussen, C. & Jordano, P. (2010). Missing and
- ⁷⁵⁷ forbidden links in mutualistic networks. *Proceedings of the Royal Society B: Biological Sciences*, 278,
- 758 725–732.
- Park, J. & Newman, M.E.J. (2004). Statistical mechanics of networks. *Physical Review E*, 70, 066117.
- Pascual, M., Dunne, J.A. & Dunne, J.A. (2006). *Ecological Networks: Linking Structure to Dynamics in Food* Webs. Oxford University Press, USA.
- Pellissier, L., Albouy, C., Bascompte, J., Farwig, N., Graham, C., Loreau, M., *et al.* (2018). Comparing species
 interaction networks along environmental gradients. *Biological Reviews*, 93, 785–800.
- Petchey, O.L., Beckerman, A.P., Riede, J.O. & Warren, P.H. (2008). Size, foraging, and food web structure.
- ⁷⁶⁵ *Proceedings of the National Academy of Sciences*, 105, 4191–4196.

- Pilosof, S., Porter, M.A., Pascual, M. & Kéfi, S. (2017). The multilayer nature of ecological networks. *Nature Ecology & Evolution*, 1, 1–9.
- Plagányi, É.E. & Butterworth, D.S. (2004). A critical look at the potential of Ecopath with ecosim to assist in
 practical fisheries management. *African Journal of Marine Science*, 26, 261–287.
- Poisot, T., Canard, E., Mouillot, D., Mouquet, N. & Gravel, D. (2012). The dissimilarity of species interaction
 networks. *Ecology Letters*, 15, 1353–1361.
- Poisot, T., Cirtwill, A.R., Cazelles, K., Gravel, D., Fortin, M.-J. & Stouffer, D.B. (2016). The structure of
 probabilistic networks. *Methods in Ecology and Evolution*, 7, 303–312.
- Poisot, T. & Gravel, D. (2014). When is an ecological network complex? Connectance drives degree
- distribution and emerging network properties. *PeerJ*, 2, e251.
- Poisot, T., Stouffer, D.B. & Gravel, D. (2015). Beyond species: Why ecological interaction networks vary
- through space and time. *Oikos*, 124, 243–251.
- Polis, G.A. (1991). Complex Trophic Interactions in Deserts: An Empirical Critique of Food-Web Theory. *The American Naturalist*, 138, 123–155.
- Pollock, L.J., Tingley, R., Morris, W.K., Golding, N., O'Hara, R.B., Parris, K.M., et al. (2014). Understanding
- co-occurrence bymodelling species simultaneously with a Joint Species DistributionModel (JSDM).
- 782 *Methods in Ecology and Evolution*, 5, 397–406.
- Pomeranz, J.P.F., Thompson, R.M., Poisot, T. & Harding, J.S. (2019). Inferring predator-prey interactions in
- ⁷⁸⁴ food webs. *Methods in Ecology and Evolution*, 10, 356–367.
- Pringle, R.M. & Hutchinson, M.C. (2020). Resolving Food-Web Structure. Annual Review of Ecology,
- 786 Evolution, and Systematics, 51, 55–80.
- Proulx, S.R., Promislow, D.E.L. & Phillips, P.C. (2005). Network thinking in ecology and evolution. Trends in
- *Ecology & Evolution*, SPECIAL ISSUE: BUMPER BOOK REVIEW, 20, 345–353.
- ⁷⁸⁹ Pulliam, H.R. (1974). On the Theory of Optimal Diets. *The American Naturalist*, 108, 59–74.
- Rooney, N. & McCann, K.S. (2012). Integrating food web diversity, structure and stability. *Trends in Ecology* & *Evolution*, 27, 40–46.

- Salim, J.A., Saraiva, A.M., Zermoglio, P.F., Agostini, K., Wolowski, M., Drucker, D.P., *et al.* (2022). Data
 standardization of plant–pollinator interactions. *GigaScience*, 11, giac043.
- Sanders, D. & van Veen, F.J.F. (2012). Indirect commensalism promotes persistence of secondary consumer
 species. *Biology Letters*, 8, 960–963.
- ⁷⁹⁶ Saravia, L.A., Marina, T.I., Kristensen, N.P., De Troch, M. & Momo, F.R. (2022). Ecological network
- ⁷⁹⁷ assembly: How the regional metaweb influences local food webs. *Journal of Animal Ecology*, 91, 630–642.
- ⁷⁹⁸ Schmeller, D.S., Julliard, R., Bellingham, P.J., Böhm, M., Brummitt, N., Chiarucci, A., *et al.* (2015). Towards a
- ⁷⁹⁹ global terrestrial species monitoring program. *Journal for Nature Conservation*, 25, 51–57.
- Shaw, J.O., Dunhill, A.M., Beckerman, A.P., Dunne, J.A. & Hull, P.M. (2024). A framework for reconstructing
- ancient food webs using functional trait data.
- ⁸⁰² Simmonds, E.G., Adjei, K.P., Cretois, B., Dickel, L., González-Gil, R., Laverick, J.H., et al. (2024).
- Recommendations for quantitative uncertainty consideration in ecology and evolution. *Trends in Ecology &*
- *Evolution*, 39, 328–337.
- Singer, M.C. & McBride, C.S. (2012). Geographic mosaics of species' association: A definition and an
 example driven by plant–insect phenological synchrony. *Ecology*, 93, 2658–2673.
- ⁸⁰⁷ Staniczenko, P.P.A., Sivasubramaniam, P., Suttle, K.B. & Pearson, R.G. (2017). Linking macroecology and
- ⁸⁰⁸ community ecology: Refining predictions of species distributions using biotic interaction networks.
- 809 *Ecology Letters*, 20, 693–707.
- Stock, M., Poisot, T., Waegeman, W. & De Baets, B. (2017). Linear filtering reveals false negatives in species
 interaction data. *SCIENTIFIC REPORTS*, 7, 45908.
- Stouffer, D.B., Rezende, E.L. & Amaral, L.A.N. (2011). The role of body mass in diet contiguity and food-web
 structure. *Journal of Animal Ecology*, 80, 632–639.
- Stouffer, D.B., Sales-Pardo, M., Sirer, M.I. & Bascompte, J. (2012). Evolutionary Conservation of Species'
 Roles in Food Webs. *Science*, 335, 1489–1492.
- Strydom, T., Bouskila, S., Banville, F., Barros, C., Caron, D., Farrell, M.J., et al. (2022). Food web
- reconstruction through phylogenetic transfer of low-rank network representation. *Methods in Ecology and*
- 818 Evolution, 13.

- 819 Strydom, T., Bouskila, S., Banville, F., Barros, C., Caron, D., Farrell, M.J., et al. (2023). Graph embedding and
- transfer learning can help predict potential species interaction networks despite data limitations. *Methods in*
- *Ecology and Evolution*, 14, 2917–2930.
- 822 Strydom, T., Catchen, M.D., Banville, F., Caron, D., Dansereau, G., Desjardins-Proulx, P., et al. (2021). A

roadmap towards predicting species interaction networks (across space and time). *Philosophical*

Transactions of the Royal Society B-Biological Sciences, 376, 20210063.

- Tylianakis, J.M., Tscharntke, T. & Lewis, O.T. (2007). Habitat modification alters the structure of tropical host-parasitoid food webs. *Nature*, 445, 202–205.
- Vázquez, D.P., Melián, C.J., Williams, N.M., Blüthgen, N., Krasnov, B.R. & Poulin, R. (2007). Species

abundance and asymmetric interaction strength in ecological networks. *Oikos*, 116, 1120–1127.

- Wagenmakers, E.-J. & Farrell, S. (2004). AIC model selection using Akaike weights. *Psychonomic Bulletin & Review*, 11, 192–196.
- Weinstein, B.G. & Graham, C.H. (2017a). On comparing traits and abundance for predicting species
 interactions with imperfect detection. *Food Webs*, 11, 17–25.
- ⁸³³ Weinstein, B.G. & Graham, C.H. (2017b). Persistent bill and corolla matching despite shifting temporal

resources in tropical hummingbird-plant interactions. *Ecology Letters*, 20, 326–335.

- Woodward, G., Benstead, J.P., Beveridge, O.S., Blanchard, J., Brey, T., Brown, L.E., et al. (2010). Chapter 2 -
- Ecological Networks in a Changing Climate. In: Advances in Ecological Research, Ecological Networks
- ed. Woodward, G.). Academic Press, pp. 71–138.
- Woodward, G., Brown, L.E., Edwards, F.K., Hudson, L.N., Milner, A.M., Reuman, D.C., et al. (2012). Climate
- change impacts in multispecies systems: Drought alters food web size structure in a field experiment.
- Philosophical Transactions of the Royal Society B: Biological Sciences, 367, 2990–2997.
- Young, J.-G., Valdovinos, F.S. & Newman, M.E.J. (2021). Reconstruction of plant–pollinator networks from
- observational data. *Nature Communications*, 12, 3911.